North American population estimates of waterbirds, vultures and hawks from migration counts in Veracruz, México

ERNESTO RUELAS INZUNZA, LAURIE J. GOODRICH and STEPHEN W. HOFFMAN

Summary
Continental-scale bird population estimates are used as a decision-support tool in conservation plans. The calculation of these estimates includes the use of density values and survey data from one or multiple sources extrapolated to the geographic scale of interest. In this paper, we use migration count data from a migration monitoring project in Veracruz, Mexico, to revise existing North American population estimates of seven species of waterbirds, New World vultures, and diurnal raptors. In two species of waterbirds, we suggest that existing estimates are low and propose a mechanism to correct these figures. In the remaining five cases, we also determine present estimates are low, but use our data to provide new continental-scale values. We discuss the importance of refining population estimates using data from a diverse suite of field methods as a means to overcome the limitations of the single-survey-based estimates.

Resumen
Las estimaciones de poblaciones de aves a escala continental son utilizadas como una herramienta de toma de decisiones en planes de conservación. Su cálculo incluye el uso de valores de densidad demográfica y datos de reconocimientos de campo de una o más fuentes que son extrapolados a la escala geográfica de interés. En este artículo utilizamos conteos de migración de un proyecto de monitoreo en Veracruz, México, para ajustar las estimaciones poblacionales para Norteamérica de siete especies de aves acuáticas, zopilotes del nuevo mundo y rapaces diurnas. En dos especies de aves acuáticas, sugerimos que las estimaciones existentes son bajas y proponemos un mecanismo para corregir estas cifras. En los restantes cinco casos, determinamos que éstas también son bajas, pero usamos nuestros datos de campo para proporcionar nuevos valores de escala continental. Discutimos la importancia de refinar la estimación de poblaciones utilizando datos de varios muestreos de campo como forma de superar las limitaciones de aquellas que se basan en una sola fuente de información.

Introduction
Estimating the true, total number of birds per unit area is one of the central problems of avian population biology and addressing this question is of utmost importance, since population estimates are decision-support tools of critical value for conservation and management. Within the context of biological and ecological inquiry, this issue has been approached from small-to-large scales, with literally dozens of techniques employed to estimate population numbers in areas that range in size roughly from $1 \times 10^2$ to $1 \times 10^6$ m$^2$ (e.g. Ralph and Scott 1981, Bibby et al. 2000).
The precision of those estimates decreases as researchers increase the geographic extent of the
study (e.g. Flather and Sauer 1996). Habitat heterogeneity and its resulting variation in density
of occupation, the scale of sampling, number of individuals that are undetected but present, and
many other sources of error, affect their precision.

At the other end of the spectrum, large-scale estimates ranging from state to continental
(roughly from $1 \times 10^8$ – $1 \times 10^{10}$ m$^2$), have received considerable attention in recent years as
benchmarks to evaluate the success of conservation measures (Fitzpatrick 2002).

The European bird conservation community initiated this approach and organized a group of
ornithologists to calculate country estimates for all species. These first results were published
15 years ago (Tucker et al. 1994), with a subsequent assessment published a decade later (Burfield
and van Bommel 2004).

Global population estimates, such as those made for waterbirds (Delany and Scott 2006),
involves the contribution of data from multiple field projects that are then summarized into
a single figure. Since Delany and Scott’s (2006) publication is the fourth iteration of this work,
they also provide calculations of population change.

In North America, the Partners in Flight coalition (PIF) released a North American Landbird
Conservation Plan (Rich et al. 2004) that contains an appendix with continental-scale population
estimates for 448 species derived predominantly from Breeding Bird Survey data (BBS). The
accuracy of these values is scored in six categories of “accuracy rating” and seven categories of
“precision.” In the geographic scope of Rich et al. (2004: 6), “North American” excludes México,
and we adopt this convention for the purpose of this paper.

These first North American estimates are a useful reference point of central importance. The
authors of them (Rich et al. 2004, Rosenberg and Blancher 2005) have acknowledged their
conservative nature and have invited the contribution of data from other sources in refining
them. Most of the reviews and discussion over the conservation plan, however, have addressed
programmatic aspects (e.g. Faaborg 2005, Vickery and Shriver 2005), and only Thogmartin et al.
(2006) have covered the topic of the population estimates, with suggestions to improve the
analyses for deriving them. Although we strongly support the idea of improving precision
through enhancements in analysis methods, we also think that an approach that incorporates
other empirical data can help us reach the goal of better figures.

In this paper, we present a case study of new North American population estimates for seven
species of waterbirds, vultures, and diurnal raptors based on our long-term research in Veracruz,
México. We recognize that our migration-count-based suggested revisions are more compelling
for species whose breeding range lies entirely north of Mexico and winter range south of
Veracruz. In three cases (waterbirds), we use our field data to indicate that existing figures are
underestimates and suggest a mechanism to produce more appropriate calculations. In the
remaining four cases (vultures and hawks), we use our migration count data to produce new
estimates. In the closing section of this paper, we discuss the importance of incorporating data
from multiple types of surveys to improve global and continental estimates, particularly (but not
exclusively) those of the PIF.

Methods
Since 1991, we have made standardized vulture, hawk, and waterbird migration counts during
(boreal) spring and autumn seasons in Veracruz, México. This region is a geographic bottleneck
for migrants. It is located at the intersection of two major mountain systems, the Sierra Madre
Oriental and the Central Volcanic Belt, which constrain the width of the Gulf Coastal Plain at
about 19° N, c. 30 km north of Veracruz City (Ruelas et al. 2000, 2005).

These migration counts are made annually from two fixed locations 11 km apart (Cardel,
19°22’N–96°22’W, elevation 29 m, and Chichicaxtle 19°24’N–96°28’W, elevation 120 m), along
an east-to-west axis (perpendicular to the front of migration). Both sites are independent samples
of the flow of birds migrating through the area, as the distance between sites does not allow
double counts, and the combined totals from these localities are used for long-term population monitoring. Because the goal of this project is monitoring rather than an attempt to quantify the total number of migrants in transit through the region, we know we are missing a proportion of the population that migrates east and west of our field stations. Most of the waterbirds regularly recorded from our migration sites rarely fly over the Gulf, whereas hawks and vultures have never been recorded flying over this large water body, making the shoreline of the Gulf of Mexico an effective barrier delimiting the eastern extent of this flyway.


**Precision of Veracruz migration counts**

Accepting our suggested revisions also implies accepting a level of precision in Veracruz migration counts. Any person who has witnessed the autumn migration in Veracruz (e.g. the c. 500 ornithologists attending the IV North American Ornithological Conference that visited our two monitoring sites in October 2006) can easily reach the conclusion that accurate counts are impossible due to the large volume of migration (Figure 1).

The method of collecting this information in the field involves a team of three field biologists per site that are randomly assigned to teams and localities. Most birds occur in large flocks that use the thermal convection of the atmosphere for cross-country flights (Pennycuick 1998). These counts were made once the thermal-climbing period had finished and birds were in their

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**Figure 1.** Field biologist Eileen Müller Guerra estimates the number of hawks in a flock past the Cardel migration monitoring site in Veracruz, México (A total of 547 within the picture frame; photo courtesy of Aldo Contreras Reyes).
straight-line gliding flights. Our field teams directly counted the number of birds when quantities were low (e.g., a few tens) and made estimations when numbers were large. Teams estimated the number of birds with the help of a hand tally-counter by counting a small group of birds, e.g. 10 individuals, and extrapolating the approximate area occupied by these 10 birds to the rest of the flock. Most commonly-used multiples in these estimations were 3, 5, 10, and 50 individuals.

This method has been applied consistently over the years, and the relatively low interannual variation in a 13-yr series of data (Table 1), suggests that the level of error of these estimates remains relatively constant (Smith et al. 2008). Variation in migration counts is lower than the variation reported among BBS routes for most raptor species (Farmer et al. 2007). We have performed some experiments using slides, bird-counting software (Wildlife Counts http://www.wildlifecounts.com/), and video, and concur with other sources of information (Erwin 1982, Hodges 1993) that our methods produce conservative estimates, particularly for large-flocking migrants. Although the magnitude of such underestimations has not been assessed quantitatively, calculating this underestimation would be a good avenue to pursue for future estimates. Estimations of lower quantities of birds, however, are more precise than those of higher magnitude.

We compared our mean, minimum, and maximum, annual autumn count of each species with existing North American population estimates from different published sources. Most water bird species estimates originate from colony counts and aerial surveys, while those of vultures and raptors are primarily from the BBS. Population estimates made by the authors listed under each species section have recently been revised by Kushlan et al. (2002), Rich et al. (2004), Delany and Scott (2006) and Partners in Flight (2007). It is important to notice that our migration count data includes both juvenile and adult individuals, as the two age classes are impossible to track separately in the field, while the estimates of all the remaining published sources reportedly include only reproductive individuals.

### Results and Discussion

American White Pelican *Pelecanus erythrorhynchos*. The global estimated population of this species is variable: 106,690 individuals (53,345 nests × 2; fide Koonz 1987); 120,000 individuals (Kushlan et al. 2002); 134,000 individuals “east of the Rocky Mountains” (King and Anderson 2005), and over 180,000 (Delany and Scott 2006: 31). The mean Veracruz count (85,679 individuals, Table 1) is lower than any of the North American estimates. However, since the majority (over 80%) of the winter range for this species lies north of our study sites in both the Gulf and the Pacific coasts (Knopf 2004), and, since these populations do not migrate through Veracruz, it is reasonable to state that Delany and Scott’s (2006) highest estimate is substantially

### Table 1. Autumn migration count totals of seven species of wading birds, vultures, and hawks from Veracruz, México, 1992–2004, and previously-published North American population estimates (see text for source data). SD, standard deviation; CV, coefficient of variation.

<table>
<thead>
<tr>
<th>Season count total (n = 13 field seasons)</th>
<th>Highest Old Estimate</th>
<th>New Estimate</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Mean</strong></td>
<td><strong>Min</strong></td>
<td><strong>Max</strong></td>
</tr>
<tr>
<td>American White Pelican</td>
<td>85,679</td>
<td>54,507</td>
</tr>
<tr>
<td>Anhinga</td>
<td>31,633</td>
<td>18,837</td>
</tr>
<tr>
<td>Wood Stork</td>
<td>36,975</td>
<td>24,915</td>
</tr>
<tr>
<td>Turkey Vulture</td>
<td>1,895,679</td>
<td>1,474,797</td>
</tr>
<tr>
<td>Mississippi Kite</td>
<td>157,199</td>
<td>32,568</td>
</tr>
<tr>
<td>Broad-winged Hawk</td>
<td>1,931,253</td>
<td>1,534,556</td>
</tr>
<tr>
<td>Swainson’s Hawk</td>
<td>812,419</td>
<td>388,916</td>
</tr>
</tbody>
</table>

*This estimate includes populations from northeastern México
low, possibly as much as 60%, assuming a homogenous winter density, and therefore this estimate needs to be revised. An appropriate method for a more accurate estimate should include colony counts and winter density estimations across the non-breeding range of the species, possibly aided with Christmas Bird Count (CBC) data. Migration counts from localities north and south of our field sites could also help refine these estimates.

**Anhinga Anhinga anhinga.** Current North American population estimates range from 20,000–34,000 individuals (10,000–17,000 pairs x 2) (Frederick and Siegel-Causey 2000 and Kushlan et al. 2002) to 51,000 individuals for the southeastern United States alone (Delany and Scott 2006: 39). A very large proportion of the population of this species overwinters in its breeding range (Frederick and Siegel-Causey 2000), and therefore is not recorded during migration in Veracruz. Our mean autumn count of 31,633 (Table 1), may record less than 40% of the individuals that belong to the migrating population (this estimate is based on the range map in Frederick and Siegel-Causey 2000), and therefore indicates that all previous population estimates are low and should also be revised. We are unable to determine how low this estimate might be, but suggest that migration counts, non-breeding season colony counts, CBC, and winter density estimations paired with non-breeding season geographic distribution data, could help obtain a more precise number.

**Wood Stork Mycteria americana.** This species is estimated to have 25,000 individuals in the United States breeding range (Coulter et al. 1999). Delany and Scott (2006: 56) estimate 28,000 individuals for the southeastern United States alone. A more recent figure of 46,000 individuals includes México and Central America (Kushlan et al. 2002).

The mean number recorded in Veracruz is higher than any of these figures (56,975, Table 1) and nearly 20% larger than the highest published estimate. These results suggest that previous population estimates for North America are conservative, but in the case of those including Mexico and Central America (Kushlan et al. 2002) these numbers are too conservative, up to 65% below, based on winter range distribution relative to our field sites, and therefore these should be revised.

Wood Storks recorded on southbound autumn migration in Veracruz include an unknown proportion of juveniles that migrate north along the Gulf coast from colonies in the wetlands of the Grijalva and Usumacinta rivers (Tabasco and Chiapas) and the region of Centla (Tabasco), and possibly from the states of Campeche and Yucatán, during their post-fledgling period from late May through July. These individuals may join southbound migrants of northern origin later in the autumn and the number of individuals recorded in autumn migration counts may not belong to populations originating exclusively in the southeastern United States, but include birds from southern México. An adequate knowledge of the post-breeding dispersal and migration ecology of this species, colony counts in the non-breeding season, and migration counts will also be helpful in producing more appropriate estimates.

**Turkey Vulture Cathartes aura.** Population estimates cover the broad and expanding range of this species throughout North America (Kirk and Mossman 1998). The United States and Canada population estimate in Rich et al. (2004) is 1,305,000 individuals (or 29% of 4,500,000 globally). Our mean annual count is higher (1,895,679, Table 1) exceeding the highest published figure by over 30%, yet, Veracruz counts detect only a portion of the North American breeding population. Much (very likely more than 50%) of the eastern population of Turkey Vultures does not migrate as far as Mexico during the winter period (Kirk and Mossman 1998).

Our field sites are not sampling a proportion of the population migrating west of our study sites (10%), since thermal-soaring obligates such as this species, are restricted to the lowlands of the coastal plain where they find abundant thermals for energy-saving migration. Our rounded estimate of the North American and northern México population migrating through Veracruz is 2,945,000 individuals. An accurate estimation of the North American population requires more information on the geography of migration of this species.

**Mississippi Kite Ictinia mississippiensis.** The North American global population estimate by Rich et al. (2004) is 190,000 individuals, very similar to the estimate of Parker (1999), who
obtained this information from our count data. This is a total migrant (breeding and wintering ranges are disjunct), whose most important migration route is the Gulf coastal plain of México, and a very high percentage of the global population of this species is recorded in our count sites. The mean number recorded in Veracruz is 157,199 (Table 1).

The number of Mississippi Kites tracked in our field sites is a sample of the total population migrating through this area. This is a facultative thermal-soaring migrant in Veracruz, and a proportion of the flight that we estimate to be 15% (and this percentage is slightly larger than in species of thermal-soaring obligates, Ruelas 2005), is missed west (inland) of our study sites; therefore, our rounded population estimate for this species is 181,000 individuals. This species is undergoing an explosive population increase, at a rate of >15% per year (Smith et al. 2008).

Broad-winged Hawk Buteo platypterus. Global population estimates of 1.7 million individuals are based on research from our study site (Goodrich et al. 1996). This estimation concurs with the one of Rich et al. (2004) of 1,728,000 individuals (96% of 1,800,000). The mean number recorded in Veracruz (1,931,255, Table 1) is slightly larger than earlier North American estimates. This species is a total migrant and, with the exception of populations overwintering in southern Florida and individuals migrating into Cuba, a very large proportion of the global population is recorded during migration in Veracruz. The Broad-winged Hawk is a thermal-soaring obligate restricted to the coastal plain and we estimate that 10% of the transient population is not sampled west of our study sites. Our rounded estimate for North America is 2,125,000 individuals.

Swainson’s Hawk Buteo swainsoni. This is another species of total migrant whose global population estimates are based on unpublished data from our research sites (Ruelas et al. 1996). During the period 1995–1996, our most conservative estimate was 448,000 (England et al. 1997), which coincides with the figure provided by Rich et al. (2004) of around 460,600 (94% of 490,000). The mean number recorded in Veracruz (1,197,850 birds, Table 1) is 56.6% larger than the previous highest population estimate.

The migration of this ‘western’ species follows a path across northern México into the Gulf coastal plain and most Swainson’s Hawks overwinter in Argentina, with the exception of populations in California, western México (Binford 1989), and Florida (England et al. 1997). The disjunct distribution of this species in the breeding and wintering area, the limited options for migration routes, and the outcome of migration counts and satellite telemetry studies suggest a very large proportion of the global population migrates through this bottleneck (e.g. Fuller et al. 1998). Swainson’s Hawk is also a thermal-soaring obligate restricted to the coastal plain and we estimate our sites are missing 10% of the total population transient through the area. Our rounded population estimate is 894,000 individuals.

Population estimates as a product of a metanalysis

The main difference between the Global Waterbird (Delany and Scott 2006), European (Burfield and van Bommel 2004), and North American (Rich et al. 2004) population assessments is that the first two use an approach that combines data from multiple types of surveys that cover the breeding and the non-breeding season, whereas the latter is based almost completely upon BBS data that most likely excludes all non-breeding adult and juvenile individuals (Rich et al. 2004: 79, Rosenberg and Blancher 2005).

The authors of the North American landbird estimates (Rich et al. 2004, Blancher et al. 2007) acknowledge the limitations of this single-survey-based approach. Including data from different surveys could enable us to derive more appropriate population estimates, since no survey is capable of generating abundance indices for all species.

This is particularly important in the case of raptors. BBS data have been shown to have limitations estimating trends and indices of abundance in species detected at rates of <1.0 birds per route (Peterjohn et al. 1995), which is the case for all the species treated in this paper.
Similarly, Farmer et al. (2007) have shown that, for most species, BBS population change estimates are of lower precision than those of migration counts.

Towards a precise conservation tool

The word of caution offered by Rosenberg and Blancher (2005: 65), that “using inaccurate population estimates to set conservation targets may lead to misdirected conservation actions” is of critical importance. One example of this is the recently recovered Peregrine Falcon (Falco peregrinus). The estimate of Rich et al. (2004) for the North America population is 276,000 (23% of 1,200,000 worldwide). However, an estimate based on nest surveys by White et al. (2002), suggests the North American population ranges between 52,000 and 62,000 individuals, less than one-quarter than reported in the PIF estimate. As we further monitor the recovery of this species and the impact of management activities on their regional populations (Millsap and Allen 2006, USFWS 2007), it becomes essential to work with the best population estimates possible.

The opportunity to have a revised database of population estimates as a conservation tool of high precision may be in front of us. The PIF population estimates are now available at the state level (Partners in Flight 2007). The PIF Science Committee use the BBS-based approach (Blancher et al. 2007) that will soon incorporate the suggestions of Thogmartin et al. (2006) at this level of detail (P. Blancher, pers. comm.). Now that these state-level, baseline data exist, the fine-grain resolution will facilitate reducing the error of these estimates and will be open to the contribution of more data from other surveys and researchers (Confer et al. 2008, Lanctot et al. 2008). An adjustment to these estimates, like the one in our case study using migration count data at a large scale, is more feasible at finer scales.

Generating these first baseline data should be considered a work in progress, rather than a final word, in the creation of this collaborative conservation tool (Lanctot et al. 2008, Farmer 2008). We believe that (1) The contribution of data derived from diverse surveys, and (2) Sampling data from individual projects that involve large proportions of the population of interest, enhances the precision and value of North American estimates.

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